Electrochemically-Driven and Green Conversion of \( \text{SO}_2 \) to \( \text{NaHSO}_4 \) in Aqueous Solution

Hong Liu\(^1,2,\ast\), Chuan Wang\(^1,2\) and Yuan Liu\(^1\)

\(^1\) Chongqing Institute of Green and Intelligent Technology, Chinese Academy of Sciences, Chongqing, 
\(^2\) School of Chemistry and Chemical engineering, Sun Yat-sen University, Guangzhou, China

1. Introduction

The world has widely resorted to fossil fuels to power the industry and everyday life. In China, above 70% of the energy is extracted from coal. Emission of \( \text{SO}_2 \) due to burning of fossil fuels, in particularly of coal, causes harmful impacts on the environment, human health, livestock, and plants \((1,2)\). Many measures have been taken to cut off the emission of \( \text{SO}_2 \) during the last generations. It can be seen that the reduction of \( \text{SO}_2 \) emission in developed countries such as the United States has been witnessed \((3)\). However, due to the aspiration of energy to drive the economic increase and sustain the expanded population, the \( \text{SO}_2 \) emission is estimated to augment sharply in the rapidly developing Asian areas, and will still pose as a worldwide environmental problem in the next 30 years \((4)\).

Basically, the \( \text{SO}_2 \) emission is reduced after the burning processes through various flue gas desulfurization (FGD) processes \((5-8)\), which serves to transform the \( \text{S}(\text{IV}) \) to \( \text{S}(\text{VI}) \) and frequently to immobilize the \( \text{SO}_2 \) waste in the form of a solid. Of them, a wet limestone FGD process \((6,9,10)\) using \( \text{CaCO}_3 \) mineral, represents over 90% of the installed desulfurization capacities in the world \((9)\), and can be chemically expressed below:

\[
\text{SO}_2 + \text{CaCO}_3 + 2\text{H}_2\text{O} + \frac{1}{2}\text{O}_2 \rightarrow \text{CaSO}_4 \cdot 2\text{H}_2\text{O} + \text{CO}_2
\]

Eq 1 illustrates that the \( \text{SO}_2 \) is transformed and immobilized in the form of \( \text{CaSO}_4 \cdot 2\text{H}_2\text{O} \), which may be commercialized as gypsum, but the incentive is little in areas including the United States \((11)\) and China because of its rich natural sources. To our knowledge, only 3% of the FGD byproduct gypsum can be reused in China. In fact, once treated improperly, the solid waste becomes a secondary pollutant, and thereby is of a great environmental concern. Meanwhile, eq 1 shows that 1 mole of \( \text{SO}_2 \) leads to 1 mole of \( \text{CO}_2 \), whose discharge and
accumulation in the atmosphere is recognized to aggravate a greenhouse effect (12,13). Actually, it is considered that most wet FGD processes have an inherent shortcoming of secondary pollution, or of high running cost if the secondary pollutant is avoided. A challenge to overcome such shortcomings still remains (14).

To meet the challenge, novel green technologies with no/less secondary pollution and with a value-added product become essential. Fan et al. have developed a process of converting the SO$_2$ to polymeric ferric sulfate, which can be employed as a common coagulant for water and wastewater treatment (12).

Electrochemical techniques, utilizing electrons as a clean reagent, exactly enjoy the sustainability. Since most wet FGD processes embrace a sub-process of electron transfer for the oxidation of S(IV) to S(VI), the electrochemical techniques appear to fitfully work there. The electrochemical cleanup of flue gas has already been tested. For example, SO$_2$ can be anodically oxidized to H$_2$SO$_4$ in aqueous solution (14-16), and regeneration of FGD agents is developed by using electrodialysis with a bipolar membrane (17,18). It should be noted that as air coexists with the SO$_2$ in flue gas, electrochemical utilization of the molecular oxygen from air to further oxide the SO$_2$ is indispensable and should be encouraged. Such a new concept, however, has not been implemented so far.

To convert the SO$_2$ to be a value-added product without secondary pollution, this study aims at developing such a novel and green process by designing a series of electrochemical reactions through a SO$_2$ absorption-and-conversion process. In the process, a few considerations in the process can benefit the attempt. (i) The cathodic reaction utilizes O$_2$ from air to scavenge the process-released H$^+$ ions, while the anodic reaction uses H$_2$O to supply H$^+$ ions. (ii) The H$^+$ scavenging benefits the SO$_2$ absorption and its further oxidation. (iii) The H$^+$ supply benefits the formation of a bisulfate. Consequently, the SO$_2$ conversion is driven electrochemically to form NaHSO$_4$ as a sulfur-containing product.

NaHSO$_4$ is a valuable chemical and widely used as an additive in manufacture of dye stuff, a soil amender in agriculture, and replacement of H$_2$SO$_4$ in industry for pH adjustment and catalytic reactions. This study focused on the chemical and sustainable fundamentals as well as the pH optimization for the SO$_2$ oxidation. The findings are expected to lay a basis of understanding this new design with potential to convert the SO$_2$ from flue gas to NaHSO$_4$ as a value-added product in a green way.

2. Chemical fundamentals of the process

In this process, the SO$_2$ is designed to be absorbed into aqueous solution with alkaline, then oxidized to sulfate, and then transformed into bisulfate. The chemical fundamentals should be clarified to understand how the process works.

Absorption of SO$_2$. In the wet FGD processes, H$^+$ ions are released upon the absorption of SO$_2$ into the aqueous solution (5):

$$SO_2 + H_2O \leftrightarrow H^+ + HSO_3^- \quad (2)$$

$$HSO_3^- \leftrightarrow H^+ + SO_3^{2-} \quad (3)$$
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Oxidation of SO$_2$ to SO$_4^{2-}$. After the absorption, the absorbed SO$_2$ is oxidized by air in aqueous solution at moderate pH from HSO$_3^-$ and SO$_3^{2-}$ to SO$_4^{2-}$ ions. This oxidation process is expressed as follows, of which eq 4 releases H$^+$ ions, but eq 5 does not (19):

\[
\text{HSO}_3^- + \frac{1}{2} \text{O}_2 \rightarrow \text{SO}_4^{2-} + \text{H}^+ \quad (4)
\]

\[
\text{SO}_3^{2-} + \frac{1}{2} \text{O}_2 \rightarrow \text{SO}_4^{2-} \quad (5)
\]

ESH. The H$^+$ ions released through eqs 2~4 need to be scavenged due to their hindrance of the SO$_2$ absorption. Otherwise, the continuous absorption of SO$_2$ will be terminated. It can be noted that while O$_2$ in air is utilized for the SO$_2$ oxidation through eqs 4 and 5, the cathodic reduction of O$_2$ can be employed to scavenge the process-released H$^+$ ions. The reactions of O$_2$ reduction through a 2-electron process at acidic and neutral/alkaline conditions are expressed in eqs 6 and 6’, respectively (20-22):

\[
\text{O}_2 + 2\text{H}^+ + 2\text{e}^- \rightarrow \text{H}_2\text{O}_2 \quad (6)
\]

\[
\text{O}_2 + 2\text{H}_2\text{O} + 2\text{e}^- \rightarrow \text{H}_2\text{O}_2 + 2\text{OH}^- \quad (6')
\]

At the same time, a side reaction co-exists with eq 6 below:

\[
2\text{H}^+ + 2\text{e}^- \rightarrow \text{H}_2 \uparrow \quad (7)
\]

It can be seen that eqs 6 and 7 consume H$^+$ ions and eq 6’ supplies OH$^-$ ions. All these reactions can be utilized to scavenge the H$^+$ ions released through eqs 2~4.

Transformation of SO$_4^{2-}$ to Bisulfate. As coupled to the cathodic reactions, an anodic reaction is shown below:

\[
\text{H}_2\text{O} \rightarrow \frac{1}{2} \text{O}_2 \uparrow + 2\text{H}^+ + 2\text{e}^- \quad (8)
\]

Under an extremely acidic condition, the H$^+$ ions in eq 8 are combined with the SO$_4^{2-}$ ions formed in eqs 4 and 5 to form a bisulfate:

\[
\text{SO}_4^{2-} + \text{H}^+ \leftrightarrow \text{bisulfate ion} \quad (9)
\]

As a result, a model experiment becomes necessary to chemically substantiate this design by disclosing the ESH effect on the SO$_2$ absorption and oxidation, and by confirming the formation of NaHSO$_4$ as a product of desulfurization.

3. Experimental section

Chemicals and Reagents. SO$_2$ gas (99.9%) was obtained from KEDI, Foshan, China. Other chemicals as analytical reagents were used as obtained. Double distilled water was used in all experiments.
**Experimental Procedure.** The model experiment to substantiate this design was performed in an experimental setup as schemed in Figure 1, which consists of two 200 mL chambers: a DS chamber with a graphite rod (ø = 6.4 mm, and L = 200 mm, Chenhua, Shanghai, China) as the cathode and a saturated calomel electrode as the reference electrode, and an SR chamber with a Pt flake (2 × 1.5 cm², Chenhua, Shanghai, China) as the anode. Both chambers were connected by a salt bridge containing saturated Na₂SO₄ solution with agar. A PS-1 potentiostat/galvanostat (Zhongfu, Beijing, China) was employed to apply a cathodic current. The solution temperature was kept by a water bath at 25.0 ± 0.5 °C and monitored by a thermometer. Air was purged onto the cathode surface by an air pump through a glass frit diffuser, and a needle valve was used to control its flow rate.

![Diagram of the experimental setup](image)

**Fig. 1.** Scheme of the experimental setup: DS chamber is desulfurization chamber, and SR chamber is sulfur-recovery chamber.

To quantify the OH⁻ ions electrochemically generated in the DS chamber and the H⁺ ions in the SR chamber, electrochemical reactions of eqs 6, 6', 7 and 8 were performed in 0.01 M Na₂SO₄ solution without SO₂ at pH₀ 6.0, in which an air flow of 100 mL min⁻¹ and different current densities of 0.10, 0.15, 0.20, 0.25, and 0.30 mA cm⁻² were applied. During the reactions, 0.01 M HCl and 0.01 M NaOH solutions were fed by a pump (Longer BTOQ-50M, Baoding, China) into the DS chamber and SR chamber, respectively, to maintain the pH at 6.0 ± 0.2. At the same time, 1.0 mL of solution sample was taken from the DS chamber for the quantification of H₂O₂ generation.

After that, two steps of the model experiment were performed in a batch mode. Step I was performed in the DS chamber, and Step II, in the SR chamber. Each experiment was performed three times and the values of experimental data in average are presented.

Prior to Step I, a start-up procedure was carried out to pre-store OH⁻ ions in the DS chamber through eqs 6', and thus a solution pH₀ ≥ 9.0 in this chamber was obtained. The start-up procedure was described below, which was performed in the setup as schemed in Figure 1. Upon application of a cathodic current density as large as 0.60 mA cm⁻² to preclude the generation of cathodic byproduct H₂O₂ (20), air with 100 mL min⁻¹ flow rate was bubbled into water in the DS chamber to allow eq 6' to occur. The coupled reaction of eq 8 occurred...
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simultaneously in the SR chamber. This start-up procedure continued till the pH increase to \( \geq 9.0 \) in the DS chamber. Then, the solution in the SR chamber was discarded, and the solution in the DS chamber containing the pre-stored OH\textsuperscript{-} ions was utilized to absorb the SO\textsubscript{2} gas in Step I of the model experiment.

Thereafter, the SO\textsubscript{2} absorption of Step I was performed without current application. Initially, nitrogen gas was bubbled into the solution to remove any oxygen, then gaseous SO\textsubscript{2} was introduced into the alkaline solution to form A-SO\textsubscript{2} till pH decreased below 7.0.

In the A-SO\textsubscript{2} oxidation of Step I, an air flow at 100 mL min\textsuperscript{-1} was purged onto the cathode placed in the A-SO\textsubscript{2} solution. And a cathodic current was applied to maintain the electrochemical reactions. The A-SO\textsubscript{2} oxidation proceeded till the solution pH recovered to neutral pH (7.0).

To optimize the pH for the A-SO\textsubscript{2} oxidation, a set of experiments was performed at 1.0 mM A-SO\textsubscript{2} concentration, and different pHs in the range of 4.0~8.0 were maintained by chemical dosing of 0.01 M NaOH solution except at pH\textsubscript{0} 8.0.

In the transformation of SO\textsubscript{4}\textsuperscript{2-} to bisulfate of Step II, the reacted solution of Step I was relocated from the DS chamber to the SR chamber. Step II proceeded under a cathodic current with solution pH decrease in the SR chamber, while it stopped upon that the pH in the DS chamber reached the pH\textsubscript{0} value at the start-up step.

In Step I, the A-SO\textsubscript{2} concentrations were monitored by taking 1.0 mL of solution samples at pre-set time intervals, and 5 \( \mu \)L of methanol was injected into the samples taken during the A-SO\textsubscript{2} oxidation to quench any possible radical reaction (23). After Step I, SO\textsubscript{4}\textsuperscript{2-} concentrations were measured. After Step II, H\textsuperscript{+}, Na\textsuperscript{+}, and SO\textsubscript{4}\textsuperscript{2-} concentrations were measured.

Notably, in actual wet FGD processes, the SO\textsubscript{2} absorption and oxidation in Step I occur concurrently. However, to understand the ESH effect on the SO\textsubscript{2} absorption and its oxidation independently, the two experiments were conducted separately. At the same time, the air content in actual FGD processes should be minimized, and thus a small rate of 100 mL min\textsuperscript{-1} was fixed without further optimization. It was believed that an alkaline condition at pH > 7.0 was beneficial to the SO\textsubscript{2} absorption, and an acidic condition at pH < 3.0 was beneficial to the NaHSO\textsubscript{4} formation, while the A-SO\textsubscript{2} oxidation relied on pH, so the pH optimization for the A-SO\textsubscript{2} oxidation was performed.

**Chemical Analysis.** Measurement of the concentrations of electrochemically-generated OH\textsuperscript{-} and H\textsuperscript{+} ions is accomplished by counting the dose of added H\textsuperscript{+} solution with a known concentration to determine the amount of OH\textsuperscript{-}, and by counting that of OH\textsuperscript{-} to determine that of H\textsuperscript{+}. The A-SO\textsubscript{2} concentration was determined in terms of S(IV) concentration using a UV-VIS spectrophotometer (TU1810, Universal Analysis, Beijing, China) according to a reported procedure (24). The H\textsubscript{2}O\textsubscript{2} concentration was determined by spectrophotometry according to a potassium titanium (IV) oxalate method (20). The SO\textsubscript{4}\textsuperscript{2-} and Na\textsuperscript{+} concentrations were measured by ion chromatography (Dionex DX-600, U.S.). The cyclic voltammetry of A-SO\textsubscript{2} solution was performed in a N\textsubscript{2}-saturated solution with 0.05 mM Na\textsubscript{2}SO\textsubscript{4} as electrolyte on a CHI work station (Chenhua, Shanghai, China) with 50 mV s\textsuperscript{-1} scanning rate. The pH was monitored by a pH meter (PB-10, Sartorius, Shanghai, China).
4. Results and discussion

Electrochemical Generation of OH⁻ and H⁺ Ions. This designed process underlined (i) that the electrochemical scavenging of process-released H⁺ ions due to eqs 2-4 would benefit the SO₂ absorption and oxidation, and (ii) that the electrochemical supply of H⁺ ions through eq 8 would realize the NaHSO₄ formation. Three cathodic reactions of eqs 6, 6', and 7 served to increase the solution pH and electrochemical generation of OH⁻ ions should be considered. Figure 2A illustrated that the OH⁻ concentration in the DS chamber increased proportionally to the reaction time and the accumulative rate of OH⁻ ions depended on the applied current density. In the meantime, H₂O₂ was generated through eq 6 and 6'. Figure 1B revealed that the H₂O₂ concentration increased against the reaction time. As paired to the cathodic reactions, anodic reaction of eq 8 occurred to supply H⁺ ions in the SR chamber. Measurements of the H⁺ ions revealed that the H⁺ concentration increased at the same rate as that of OH⁻ ions in the DS chamber (not shown here).

The electrochemically-generated OH⁻ ions in the DS chamber and H⁺ ions in the SR chamber carried electrons, which must be balanced electrically to keep the electrical neutralization of solution in each chamber. An analysis of electron balance will be described later.

ESH Effect on SO₂ Absorption. The above electrochemically-generated OH⁻ ions could be utilized to scavenge the absorption-released H⁺ ions through eqs 2 and 3. ESH has two functions. One was the pre-storage of OH⁻ ions through eq 6', which later served to scavenge the SO₂-absorption-released H⁺ ions. The other was the in situ scavenging of the process-released H⁺ ions through eqs 6 and 7 or in situ supplying OH⁻ ions through eq 6', which also served to scavenge the A-SO₂-oxidation-released H⁺ ions.

To disclose the ESH effect on the SO₂ absorption, one SO₂ absorption experiment in Step I of the model experiment was performed at pH₀ 9.0 with pre-stored OH⁻ ions, and then two additional SO₂ absorption experiments without pre-stored OH⁻ ions were performed at pH₀ 5.0 and 7.0, respectively. Figure 3 revealed that under the three pH conditions, the A-SO₂
increased first quickly, then slowly to a plateau. Comparatively, at pH 9.0, the A-SO$_2$ concentration increased most rapidly and ended at the highest level. The acceleration of SO$_2$ absorption at pH 9.0 was caused by the lifted pH, which was realized by pre-storing the OH$^-$ ions electrochemically. Obviously, the pre-stored OH$^-$ ions served to scavenge the SO$_2$-absorption-released H$^+$ ions as a means of ESH. Thus, the SO$_2$ absorption was accelerated by the ESH.

Fig. 3. Buildup of A-SO$_2$ concentration in the SO$_2$ absorption at solution pH 5.0, 7.0, and 9.0.

**ESH Effect on A-SO$_2$ Oxidation.** The oxidation-released H$^+$ ions were in situ scavenged through eqs 6, 6’ or 7. To disclose the ESH effect on the A-SO$_2$ oxidation, two sets of SO$_2$ oxidation experiments in Step I were carried out. In the first set with ESH, a current density at 0.20 mA cm$^{-2}$ was applied to maintain the electrochemical reactions, while in the second set without SH, no current density was applied.

Figure 4 showed that with ESH, 100% and 95.8% of A-SO$_2$ disappearances were achieved at 30 min for the 1.0 mM A-SO$_2$ and 1.5 mM A-SO$_2$ solutions, respectively. Following the oxidation reaction, SO$_4^{2-}$ concentrations were detected, and the results as listed in Table 1 indicated that 95.0% and 88.0% of the A-SO$_2$ were converted to SO$_4^{2-}$ ions. By contrast without SH, only 70.7% of the 1.0 mM A-SO$_2$ and 60.9% of the 1.5 mM A-SO$_2$ disappeared after 30 min, and the SO$_4^{2-}$ concentrations in the reacted solution were significantly lower than those with ESH (Table 1). From these it could be understood that the A-SO$_2$ oxidation with ESH proceeded more rapidly than that without SH, since ESH was beneficial to the conversion of A-SO$_2$ oxidation to SO$_4^{2-}$. 

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Fig. 4. Temporal disappearance of A-SO$_2$ concentration in the A-SO$_2$ oxidation at pH$_0$ 6.0, with ESH at 0.20 mA cm$^{-2}$ current density for 1.0 mM A-SO$_2$ and 0.25 mA cm$^{-2}$ current density for 1.5 mM A-SO$_2$, with CSH, and without SH.

Table 1. SO$_4^{2-}$ concentrations after the A-SO$_2$ oxidation at pH$_0$ 6.0, with ESH at 0.20 mA cm$^{-2}$ current density for 1.0 mM A-SO$_2$ and 0.25 mA cm$^{-2}$ current density for 1.5 mM A-SO$_2$, with CSH, and without SH.

<table>
<thead>
<tr>
<th>A-SO$_2$ (mM)</th>
<th>SO$_4^{2-}$ with ESH (mM)</th>
<th>SO$_4^{2-}$ with CSH (mM)</th>
<th>SO$_4^{2-}$ without SH (mM)</th>
</tr>
</thead>
<tbody>
<tr>
<td>1.0</td>
<td>0.98</td>
<td>0.78</td>
<td>0.66</td>
</tr>
<tr>
<td>1.5</td>
<td>1.32</td>
<td>1.15</td>
<td>0.97</td>
</tr>
</tbody>
</table>

To further disclose the ESH effect, a set of experiments in 1.0 mM and 1.5 mM A-SO$_2$ solutions was carried out with CSH, and the results are added in Figure 4. Clearly, the A-SO$_2$ oxidations with CSH proceeded more rapidly than those without SH, but more slowly than those with ESH. Therefore, it was further confirmed that the SH benefited the A-SO$_2$ oxidation, while the ESH was more effective than the CSH.

Moreover, the results in Table 1 showed that more SO$_4^{2-}$ ions were obtained in the reacted solution with ESH than those with CSH. It was believed that the H$_2$O$_2$ produced through eq 6 or 6' could enhance the SO$_2$ oxidation as an oxidizing reagent (27,28). After the reaction, no H$_2$O$_2$ residue left as impurity in the final sulfur-containing product. Therefore, the ESH benefited the A-SO$_2$ oxidation with two advantages of (i) scavenging the absorption- and oxidation-released H$^+$ ions, and (ii) simultaneously generating H$_2$O$_2$ to facilitate the conversion of A-SO$_2$ to SO$_4^{2-}$. 

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Transformation of Na$_2$SO$_4$ to NaHSO$_4$. Beyond the utilization of cathodic reaction to scavenge the H$^+$ ions released in eqs 2~4, this new process utilized an anodic reaction (eq 8) to supply H$^+$ ions which combined the SO$_4^{2-}$ to form bisulfate. In both Steps I and II, the H$^+$ ions were produced in the SR chamber. Step II was performed by relocating the reacted solution of Step I from the DS chamber to the SR chamber where eq 8 occurred. Thus after this step, three ions of Na$^+$, H$^+$, and SO$_4^{2-}$ presented in the SR chamber, and the analysis of mass balance of electrons as would be shown later suggested that a mixture of the three ions might result in the formation of NaHSO$_4$ through eq 9.

To confirm the NaHSO$_4$ formation, after the two-step model experiments with 1.0 mM and 1.5 mM A-SO$_2$ concentrations, the concentrations of the three ions in the SR chamber were measured. The results as shown in Figure 5 demonstrated that a mass balance of Na$^+$, H$^+$ and SO$_4^{2-}$ ions was approximately 1:1:1, which ensured the NaHSO$_4$ formation.

![Figure 5](image)

**Fig. 5.** Ion concentrations measured in the SR chamber after the model experiment.

Optimization of pH for A-SO$_2$ Oxidation with ESH. Figure 6A demonstrated that pH 5.0~6.0 was optimal for the A-SO$_2$ oxidation, which was in good agreement with others’ results of optimal pH 6.0 kept by addition of OH$^-$ ions in the SO$_2$ oxidation by air during a seawater FGD process (24). Figure 6B showed the records of cyclic voltammetry in A-SO$_2$ solution, which confirmed that the optimal pH for the SO$_2$ oxidation was 5.0~6.0.

Figure 6B illustrates the cyclic voltammetry recorded in the 1.0 mM A-SO$_2$ solution, and the peaks at 0.12~0.15 V and 0.81~0.83 V were associated with the HSO$_3^-$ oxidation and the SO$_3^{2-}$ oxidation, respectively. On the other hand, the A-SO$_2$ solution consists of two major species of HSO$_3^-$ and SO$_3^{2-}$ in the pH range of 4.0~9.5. It has been documented that the HSO$_3^-$ species occupies 100%, 92%, 38% and 10% of A-SO$_2$ at pH 5.0, 6.0, 7.0 and 8.0, respectively (27).
Clearly, the peaks associated with the HSO$_3^-$ oxidation were not pronounced at pH 7.0 and 8.0, while the peaks were mature at pH 5.0 and 6.0. On the contrary, the peaks associated with the SO$_3^{2-}$ oxidation were mature at pH 7.0 and 8.0, while no peak was observed at 5.0 and 6.0 since the SO$_3^{2-}$ species were only 8% and 0%, respectively. These results implied that at optimal pH 5.0~6.0, the HSO$_3^-$ predominated in the A-SO$_2$. Moreover, the HSO$_3^-$ oxidation proceeded more rapidly at this pH range than at pHs beyond this range. Thus, the optimal pH for the A-SO$_2$ oxidation was pH 5.0~6.0.

Fig. 6. Dependence of 1.0 mM A-SO$_2$ disappearance after 30 min oxidation on the solution pH (A), and cyclic voltammetry recorded in 1.0 mM A-SO$_2$ solution (B).

In real application, the current density should be adjusted to maintain the optimal range of pH 5.0~6.0 for the A-SO$_2$ oxidation. Consequently, a set of 1.0 mM A-SO$_2$ oxidation experiments was performed at pH 5.0, 6.0, 7.0, and 8.0. It was found that when the current densities were adjusted to 0.20, 0.25, and 0.30 mA cm$^{-2}$, the optimal pH 5.0~6.0 could be maintained for pH 7.0, 6.0, and 5.0, respectively. Evidently, the current density to maintain the optimal pH increased with the pH decrease of A-SO$_2$ solution. The lower pH meant more H$^+$ ions in the solution, so the higher current density was required to scavenge the H$^+$ ions. At pH 8.0, the SO$_3^{2-}$ species predominated in the A-SO$_2$ solution (27), and no pH decrease was observed in the A-SO$_2$ oxidation by air as indicated by eq 5. Thus no application of cathodic current to generate OH$^-$ ions was required.

Analysis of Electron Balance in DS Chamber and SR Chamber. In this study, OH$^-$ ions were generated through eq 6’ in the DS chamber, and H$^+$ ions were generated through eq 8 in the SR chamber. The electrons carried by the OH$^-$, H$^+$, and other emerged ions must be balanced electrically to keep the electrical neutralization of solution in either chamber. Table 2 lists the ion species after each step in the DS chamber and SR chamber. From the ion species, the analysis of electron balance could be made.
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Table 2. Ion species after each step in the DS chamber and SR chamber.

<table>
<thead>
<tr>
<th>DS chamber</th>
<th>SR chamber</th>
</tr>
</thead>
<tbody>
<tr>
<td>Step I</td>
<td>Step I</td>
</tr>
<tr>
<td>HSO$_3^-$</td>
<td>H$^+$ from eq 8</td>
</tr>
<tr>
<td>Na$^+$ (pre-stored)</td>
<td>SO$_4^{2-}$ from salt bridge</td>
</tr>
<tr>
<td>SO$_4^{2-}$ from eq 4</td>
<td></td>
</tr>
<tr>
<td>Na$^+$ from salt bridge</td>
<td></td>
</tr>
<tr>
<td>OH$^-$ from eq 6’</td>
<td>H$^+$ and SO$_4^{2-}$ coming from Step I</td>
</tr>
<tr>
<td>---</td>
<td>Na$^+$ and SO$_4^{2-}$ relocated from DS chamber of Step I, H$^+$ from eq 8, SO$_4^{2-}$ from salt bridge</td>
</tr>
</tbody>
</table>

In the DS chamber, Step I, the absorption-released H$^+$ ions through eqs 2 and 3 were scavenged by the pre-stored OH$^-$ ions. After Step I, HSO$_3^-$ ions predominated in the A-SO$_2$ solution at pH 5.0~6.0 (27). Clearly, the HSO$_3^-$ ions were electrically balanced by the pre-stored Na$^+$ ions.

In the DS chamber, Step I, the oxidation-released H$^+$ ions through eq 4 were scavenged by the OH$^-$ ions that were in situ generated through eq 6, 6’, or 7. The A-SO$_2$ oxidation through eq 4 generated SO$_4^{2-}$ ions as product of Step I. Concurrently, Na$^+$ ions were released from the salt bridge to balance the SO$_4^{2-}$ ions electrically. In the SR chamber, Step I, accompanied generation of H$^+$ ions occurred through eq 8. The H$^+$ ions were balanced electrically by the SO$_4^{2-}$ ions released from the salt bridge (Figure 1).

In the SR chamber, Step II, the H$^+$ and SO$_4^{2-}$ ions from Step I that were electrically balanced mutually from Step I stayed there. And the mutually balanced Na$^+$ and SO$_4^{2-}$ ions from Step I were relocated to this chamber. At the same time, H$^+$ ions continued to be generated through eq 8. Accordingly, SO$_4^{2-}$ ions continued to release from the salt bridge to balance the H$^+$ ions. Therefore, the electron balance resulted in a mixture of H$_2$SO$_4$ and Na$_2$SO$_4$ in the SR chamber after Step II. Since a reaction of Na$_2$SO$_4$ and H$_2$SO$_4$ is often adopted to manufacture NaHSO$_4$ in industry, NaHSO$_4$ might be obtained as a product in the SR chamber after Step II.

Additionally in the DS chamber, Step II, accompanied generation of OH$^-$ ions occurred through eq 6’. The OH$^-$ ions were balanced electrically by the Na$^+$ released from the salt bridge (see Figure 1).

Process of Electrochemically-driven Conversion of SO$_2$ to NaHSO$_4$. Up to now, this process design has been chemically substantiated, and the oxidation reaction of A-SO$_2$ can be optimized. Accordingly, a process of the SO$_2$ conversion to NaHSO$_4$ was schemed in Figure 5 to show the mass flows of SO$_2$, OH$^-$ and H$^+$ ions, while those of Na$^+$ and SO$_4^{2-}$ ions are presented in Figure 8. Figures 7 and 8 illustrate that the SO$_2$ gas is converted to NaHSO$_4$ through two stages: (i) SO$_2$ absorption plus oxidation to Na$_2$SO$_4$ in the DS chamber, and (ii) transformation of Na$_2$SO$_4$ to NaHSO$_4$ in the SR chamber.

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Fig. 7. Mass flow of $\text{SO}_2$, $\text{OH}^-$, and $\text{H}^+$ ions in the process of $\text{SO}_2$ conversion to $\text{NaHSO}_4$ in aqueous solution: the number before the species designates their mole mass.

Fig. 8. Mass flow of $\text{Na}^+$ (A) and $\text{SO}_4^{2-}$ (B) ions in the proposed process of $\text{SO}_2$ conversion to $\text{NaHSO}_4$ in aqueous solution: the number before the species designates their mole mass; chambers in left hand are DS chamber, and in right hand, SR chamber.
Figure 7 illustrates that in the first stage, the SO$_2$ gas with the co-existing air is introduced into the aqueous solution where the reactions through eqs 2~7 occur to form Na$_2$SO$_4$. The SO$_2$ absorption occurs effectively at high pH, and the A-SO$_2$ oxidation occurs rapidly at pH 5.0~6.0 but quite slowly at pH 8.0. Thus, although occurring concurrently, the SO$_2$ absorption predominates first at pH above 6.0, and then the A-SO$_2$ oxidation becomes the main reaction at pH 5.0~6.0. In the second stage, the solution containing Na$_2$SO$_4$ is relocated from the DS chamber to the SR chamber, and then the Na$_2$SO$_4$ is transformed into NaHSO$_4$ (eq 9) under an acidic condition through eq 8.

Accompanied with the A-SO$_2$ oxidation in the DS chamber, H$^+$ ions are generated (eq 8) in the SR chamber for the subsequent formation of NaHSO$_4$. Accompanied with the NaHSO$_4$ transformation in the SR chamber, OH$^-$ ions are generated (eq 6') in the DS chamber and pre-stored for the operation of next run.

To practice this new process, some findings from this study may be summarized as tips: (i) alkaline condition at pH > 7.0 in the DS chamber is beneficial to the SO$_2$ absorption, which can be achieved by the electrochemical pre-storage of OH$^-$ ions through eq 6', (ii) electrochemical generation of H$_2$O$_2$ (eqs 6 and 6') in the DS chamber can be designed as a concurrent reaction with the A-SO$_2$ oxidation reaction by air, since the H$_2$O$_2$ can accelerate the A-SO$_2$ oxidation reaction significantly, and (iii) a suitable current density shall be adjusted to maintain the optimal pH 5.0~6.0 of the A-SO$_2$ oxidation reaction.

Thus by combination of the mass flow in Figure 8, an overall reaction can be expressed below to illustrate the SO$_2$ conversion into NaHSO$_4$ under the electrical driving force:

$$\text{SO}_2 + \frac{1}{2} \text{O}_2 + \text{H}_2\text{O} + \text{Na}_2\text{SO}_4 \xrightarrow{\text{electricity}} 2\text{NaHSO}_4$$  \hspace{1cm} (10)

Eq 10 indicates that from a thermodynamic point of view, the electrochemical reactions of eqs 6, 6', 7 and 8 may not be essential, but from an engineering point of view, they are critical. The cathodic reactions scavenge the absorption- and oxidation-released H$^+$ ions to drive the SO$_2$ absorption and oxidation, and the anodic reaction provides H$^+$ ions to drive the NaHSO$_4$ formation. Therefore, the electrochemical reactions ensure the process of SO$_2$ conversion to NaHSO$_4$ through eq 10.

**Sustainability Evaluation of The Process.** Green chemistry has a set of principles to minimize pollution as far as possible in chemical processes, including achieving high value of atom economy in the synthetic process, eliminating toxicity to human health and the environment, minimizing energy consumption, utilizing clean materials, and so on (28,29). Although the aspiration of green chemistry is preferably realized in the manufacturing processes rather than the subsequent cleanup of effluent, the principles are still applicable in the wet FGD process. In this study, sustainability of this new process was evaluated from the perspective of green chemistry.

The atom economy is expressed by $AU$, which is defined by a ratio of mole mass between the desired product and the reactant(s) as shown below:

$$AU = \frac{\text{mole mass in desired product}}{\text{mole mass in reactant(s)}} \times 100\%$$  \hspace{1cm} (11)
The AU values of eqs 1~10 are listed in Table 3. Clearly, eqs 5, 6, 9, and 10 received 100% of AU. Eqs 2, 3, and 4 received AU as high as 98.8%, 98.8%, and 90.0%, respectively. Only eqs 7 and 8 received 50.0% and 11.1%, respectively.

<table>
<thead>
<tr>
<th>eq</th>
<th>AU (%)</th>
<th>desired product</th>
<th>byproduct</th>
</tr>
</thead>
<tbody>
<tr>
<td>1</td>
<td>75.5</td>
<td>CaSO₄ 2H₂O</td>
<td>CO₂</td>
</tr>
<tr>
<td>2</td>
<td>98.8</td>
<td>HSO₃⁻</td>
<td>H⁺</td>
</tr>
<tr>
<td>3</td>
<td>98.8</td>
<td>SO₃⁻</td>
<td>H⁺</td>
</tr>
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<td>4</td>
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<td>H⁺</td>
</tr>
<tr>
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<td>100</td>
<td>no</td>
<td>no</td>
</tr>
<tr>
<td>6</td>
<td>100</td>
<td>H₂O₂</td>
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</tr>
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<td>7</td>
<td>50.0</td>
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<tr>
<td>6'</td>
<td>----</td>
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</tr>
<tr>
<td>8</td>
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<td>O₂</td>
</tr>
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</tr>
<tr>
<td>10</td>
<td>100</td>
<td>NaHSO₄</td>
<td>no</td>
</tr>
</tbody>
</table>

Table 3. Values of Atom Utilization (AU), desired products, and byproducts in eqs 1~10.

In addition to the individual AU values of single reaction, the byproducts of eqs 1~10 are listed in Table 3, and their fates and environmental impacts are discussed as follows. The H⁺ ions as byproduct in eqs 2~4 are scavenged to form H₂O, and the H₂O can be reused through eq 8 to generate H⁺ ions that entirely end up in the NaHSO₄ through eq 9. The H₂O₂ in eq 6 or 6' is reused to enhance the A-SO₂ oxidation and also enters into the NaHSO₄. The gaseous byproducts of H₂ in eq 6' and O₂ in eq 8 escape into the environment, whereas it is not hazardous to human health and the environment, while some safety measure should be taken to deal with the H₂. However, all other atoms are kept in the final product through reusing the byproducts in the same setup.

Therefore, it can be seen that except the portion of oxygen and hydrogen atoms that released as gases through eqs 7 and 8, respectively, all other atoms are remained in the final product. This process in essence utilizes the raw materials of Na₂SO₄, O₂ in ambient air, and water which are all environmentally clean. As a result, secondary pollution can be avoided.

The electricity consumption (E) in terms of kWh through eqs 6, 6', 7 and 8 is calculated in light of the two-step model experiment.

In the A-SO₂ oxidation of Step I, electricity was consumed through the cathodic reactions of eqs 6, 6', and 7 to scavenge the H⁺ ions released in eqs 2~4, and its consumption was calculated by:

\[
E = \frac{\bar{U} \cdot I \cdot t \cdot 10^{-3}}{64 \cdot (C_0 - C_t) \cdot 10^{-3} \cdot S} \quad (S1)
\]

where \(\bar{U}\) and \(I\) were the voltage in average and current that kept constant, respectively, the \(C_0\) and \(C_t\) were the A-SO₂ concentrations before and after the oxidation reaction, respectively, and \(S\) was the volume.
In this study, $C_0 = 1.0 \times 10^{-3}$ M, $C_t = 0.02 \times 10^{-3}$ M, $I = 8.0 \times 10^{-3}$ A, $U = 4.0$ V, $t = 0.5$ h, and $S = 0.20$ L; $C_0 = 1.5 \times 10^{-3}$ M, $C_t = 0.18 \times 10^{-3}$ M, $I = 10.0 \times 10^{-3}$ A, $U = 4.3$ V, $t = 0.5$ h, and $S = 0.20$ L. Thus, in both cases, 1.28 kWh of electricity was consumed in the oxidative conversion of every kg of A-SO$_2$ to Na$_2$SO$_4$ in Step I.

Experimental results showed that the H$^+$ ions scavenged by eqs 6, 6', and 7 in the A-SO$_2$ oxidation of Step I were equivalent to the H$^+$ ions electrochemically generated through eq 8 in Step II. Thus, the $E$ values of the A-SO$_2$ oxidation and Step II were equal, and doubled electricity of A-SO$_2$ oxidation, i.e. 2.56 kWh was consumed to convert every kg A-SO$_2$ to NaHSO$_4$. Figure 7 demonstrates that the OH$^-$ ions were pre-stored through eq 6' in Step II. Obviously, the electricity consumption to pre-store the OH$^-$ ions for Step I was included in Step II. Therefore, a total electricity of 2.56 kWh was needed for the overall conversion of every kg of gaseous SO$_2$ to NaHSO$_4$ in aqueous solution, which was only slightly higher than the electricity consumption between 1.8 and 2.4 kWh for the anodic oxidation conversion of one kg of SO$_2$ into H$_2$SO$_4$ in aqueous solution (14). In comparison, the electricity consumption by the electrochemical reactions of this new process seems quite competitive.

Also important, the final product of NaHSO$_4$ has an added value. Eq 10 indicates that 2.06 kg of Na$_2$SO$_4$ needs to be consumed to convert one kg of SO$_2$ to attain 3.75 kg of NaHSO$_4$ with 82% increase in mass. It is learned from the current market that NaHSO$_4$ has an approximately quadruple commercial value of Na$_2$SO$_4$. In addition, the alkaline and acid demanded by this new process are provided on site electrochemically instead of addition of chemicals. Considering that the alkaline and acid are manufactured at the price of electricity, we can presume that this approach appears clean and cheap.

Therefore, this new process can fully comply with the principles of green chemistry and shows promising feasibility. If it is flexibly applied in a wet FGD process for SO$_2$ removal, it could be an environmentally-sustainable technique. In fact, few of environmental processes, which serve to decompose the pollutants, have high value of atom economy. Fortunately, all the atoms from the raw materials in this process end up in the product of NaHSO$_4$. During the process, some intermediates are involved, while they are reusable and environmentally benign.

**Further Investigations.** Before this process is practiced for reduction of SO$_2$ in flue gas, more investigations remain. First, effect of CO$_2$ which abounds in the flue gas should be precluded. We will demonstrate elsewhere that the CO$_2$ can be separated advisably from the SO$_2$ in this process, and then the CO$_2$ can be further captured and recovered electrochemically in the manner of synchronous supply of alkaline and acid. Second, a side reaction that accompanies the cathodic reaction of O$_2$ reduction is the H$_2$ evolution (eq 7). A coupling of this reaction with the anodic oxidation of H$_2$O (eq 8) shows the well-know reaction of H$_2$O electrolysis. While eq 7 also outputs OH$,^-$ the reaction of H$_2$O electrolysis consumes extra energy. Thus, the H$_2$O electrolysis must be avoided by a careful operation of the electrochemical reactor and selection of cathode on which the H$_2$ evolution can be inhibited. Third, eq 6 has a byproduct of H$_2$O$_2$, which serves to accelerate the S(IV) oxidation. Under some circumstances such as on Pt modified carbon electrode, the O$_2$ reduction proceeded in a 4-electron pathway to solely output OH$^-$ free of H$_2$O$_2$. In this case,
the S(IV) oxidation is accomplished by the air oxidation while the overall reaction of NaHSO₄ production remains. Since the air oxidation of S(IV) is fast to some extent, and thus in practice, whether the H₂O₂ is essential needs further scrutiny. Forth, a salt bridge is employed in this study to spatially separate the cathodic and anodic chambers. In real application, a membrane (17,18) that is commercially available can be employed. Anyway, this new process is promising as an alternative FGD process that immobilizes the SO₂ waste in the form of non-calcium product by means of a cheap and non-toxic material, and thereby avoids the concern over any secondary pollution (30).

5. Nomenclature
A-SO₂ = absorbed SO₂ in aqueous solution
AU = atom utilization in %
CSH = chemical scavenging of H⁺ ions through addition of NaOH solution
DS = desulfurization chamber, as cathodic chamber
ESH = electrochemical scavenging of H⁺ ions through eqs 6, 6’ or 7
FGD = flue gas desulfurization
SH = scavenging of H⁺ ions
SR = sulfur-recovery chamber, as anodic chamber

6. Acknowledgements
This work was supported by Natural Science Foundation of China (Project No: 50978260, 21077136).

7. References

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Electrochemically-Driven and Green Conversion of SO$_2$ to NaHSO$_4$ in Aqueous Solution


Green chemistry is chemistry for the environment. It is really a philosophy and way of thinking that can help chemistry in research and production to develop more eco-friendly solutions. Green chemistry is considered an essential piece of a comprehensive program to protect human health and the environment. In its essence, green chemistry is a science-based non-regulatory and economically driven approach to achieving the goals of environmental protection and sustainable development. Combining the technological progress with environmental safety is one of the key challenges of the millennium. In this context, this book describes the environmentally benign approaches for the industries as well as chemical laboratories. In order to provide an insight into step change technologies, this book was edited by green organic chemists.

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